DIAGNOSTIC/FEASIBILITY STUDY REPORT BURKE LAKE GREGORY COUNTY, SOUTH DAKOTA

OFFICE OF WATER RESOURCES MANAGEMENT SOUTH DAKOTA DEPARTMENT OF ENVIRONMENT AND NATURAL RESOURCES

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Prepared by

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SUMMARY

Burke Lake is a man-made lake that has been experiencing nuisance growth of bluegreen algae and aquatic macrophytes, odor problems and fish kills. This study was initiated to provide a general assessment of the water quality of Burke Lake, identify sources of lake water quality degradation, and to propose restoration alternatives to improve the lake's water quality.

In March 1989, the South Dakota Department of Environment and Natural Resources began a Diagnostic/Feasibility Study on Burke Lake under a contracted agreement with the City of Burke and with the participation of the South Dakota Department of Game, Fish and Parks.

The study consisted of water quality monitoring of the lake and it's tributaries to evaluate watershed impacts on the lake. Limited groundwater monitoring was also conducted to evaluate potential impacts from an abandoned city landfill in the lake's watershed. In addition, sedimentation and algal surveys were conducted. Most field monitoring and surveys were completed by July, 1990. Groundwater monitoring was completed in October, 1990.

Results of the study indicated that tributaries to the lake repeatedly experienced contamination with high levels of fecal coliform bacteria and nutrients. Contaminants carried to the lake from the tributaries are the primary cause of the lake water quality degradation.

Watershed sources of the tributary loads of bacteria and nutrients included a dairy feedlot operation and animal pastures upgradient of the lake.

Data from this study and from an earlier groundwater study conducted by the South Dakota Geological Survey for the City of Burke (Hammond, 1990) indicate that nutrient contamination of the groundwater in the watershed is occurring. However, the nutrient levels found in groundwater were not substantial enough to account for the gross nutrient contamination of the tributaries.

The data suggest that a potential source of groundwater nutrient contamination exists within or close to the southeastern city limits. The abandoned city landfill did not appear to be the source of nutrients found in groundwater downgradient from the landfill. A more extensive groundwater investigation would be needed to thoroughly assess groundwater quality in the watershed, sources of contamination, and impacts on the lake.

Recommendations for lake restoration included 1) implementing appropriate engineering and conservation practices to control surface water pollution from dairy and pasture practices in the watershed, 2) conducting education activities directed toward all watershed land owners to reduce nutrient and sediment contributions to water resources, 3) undertaking a more extensive groundwater quality and lake impact study, 4) repairing the dam, 5) removing nutrient laden sediments from the lake, 6) removing nutrients from the lake water, 7) creating supplemental fisheries habitat and 8) designing and implementing an ongoing lake management/protection plan.

Table of Contents

그렇다는 맛이 되는 이는 이 사이를 하고 있다고 얼굴했다면 하는 것이라고 하는 것이다면 없다.	1 agc
Introduction	1
Study Area Description	1
Background/Historical Information	1
Water Quality Standards	3
Methods and Materials	4
Lake and Tributary Sampling	4
Groundwater Sampling	6
Analytical Methods	6
Sediment Sampling and Survey	7
Biological Sampling	7
Watershed Inspection	7
Results and Discussion	8
In-Lake Water Quality	8
Trophic Condition	8
Algae	9
Phosphorus	9
Nitrogen	9
Fecal Coliform Bacteria	11
Fisheries Impacts	11
Tributary Water Quality	11
Groundwater Investigation	12
Sediment Survey	13
Elutriate Analysis	13
Watershed Inspection	13
Dam Inspection	13

Conclusions	16
Restoration Alternatives	17
Watershed Restoration	17
In-Lake Restoration	17
Recommendations	20
References	22
APPENDIX A. Water Quality Data	23
APPENDIX B. Lake Basin Cross Sections	29
APPENDIX C. Elutriate Sample Data	41
APPENDIX D. Watershed Photographs	43

List of Tables

Table		Page
1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	Burke Lake Water Quality Standards	3
2	Sampling Period and Number of Samples	4
3	Water Quality Parameters	6
4	Analytical Methods for Physical and Chemical Parameters	7
5	In-Lake Sampling Data Mean Values	8
6	Results of October, 1990 Groundwater and Tributary Sampling	12
	List of Figures	
	기는 이 그 교육을 하는 이 사람들이 하는 사람들이 함께 되었다. 그는 사람들이 되었다. 그 사람들이 가장 하는 것이 되는 것이 나왔습니다. 그런 사람들이 되었다.	
Figure		Page
1	Burke Lake Study Site Location Map	2
2	Burke Lake Sampling Site Location	5
3	Algal Cell Counts for Burke Lake	10
4	Burke Lake Bathymetric Map	14
5	Burke Lake Sediment thickness Contour Map	15
	그런 그는 어느 그는 그래? 한 학생들은 그는 학생들 점점을 하는 것들이 하는 것이 되었다. 그렇게 되었다는 그를 가는 그를 가는 그를 가는 것이 되었다. 그는 그를 가는 그를 가는 것이 없다.	

BURKE LAKE DIAGNOSTIC/FEASIBILITY REPORT

INTRODUCTION

Burke Lake has been experiencing degraded conditions which impair the lake's designated beneficial uses which include sport fishing and swimming. The lake experiences dense algal blooms, nuisance growth of aquatic macrophytes, odor problems and fish kills. In addition, there have been reports of swimmers experiencing ear and eye infections, and itching.

In response to local inquiries for assistance, the State of South Dakota Department of Environment and Natural Resources (DENR) working in cooperation with the the local community, designed a Diagnostic/Feasibility (D/F) study for the lake. The study was implemented to identify and assess the current condition of the lake, determine water quality problems and pollution sources, and develop lake restoration alternatives.

Comprehensive water quality monitoring of tributary and in-lake waters was conducted during the period from March 28, 1989 through June 26, 1990. A sedimentation survey was conducted in February, 1989 and algal quantifications were conducted from March, 1989 through March, 1990. In addition, a groundwater investigation was conducted during October, 1990 to evaluate potential impacts from an abandoned landfill located within the watershed.

Data summaries, discussions, and management recommendations are presented in this report.

Study Area Description

Burke Lake is located within Gregory County in the extreme south central area of South Dakota. The lake has a surface area of approximately 36 acres and a watershed which encompasses 1568 acres (Figure 1).

The Burke Lake watershed consists primarily of shortgrass prairie and rangeland with a small portion being cultivated. A dairy operation is located along the southwest shore of the lake. The soils of the entire watershed are of the Anselmo-Holt-Tassel association which identifies them as welldrained loamy soils on the uplands underlain by sand or sand and gravel.

Average annual precipitation within the watershed is nearly 22 inches. Burke Lake is the principle receiving water body in the watershed and it discharges into Coon Creek which flows east to the Missouri River.

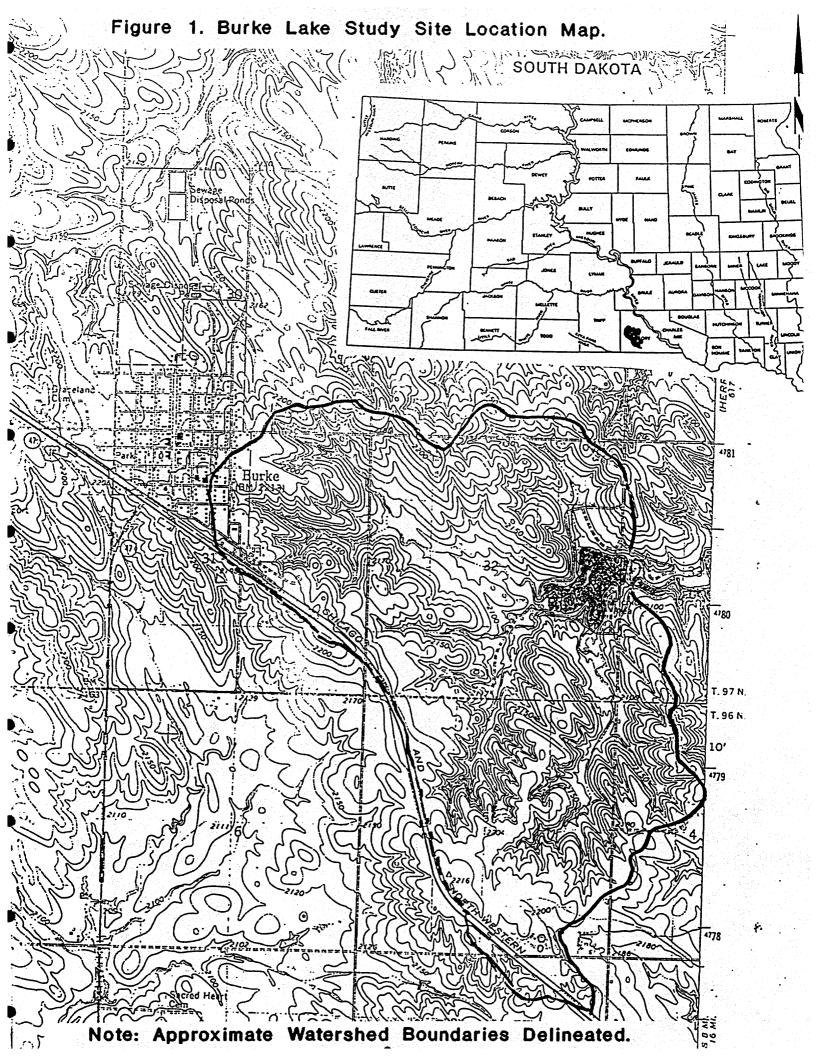
Burke Lake provides angling, boating, picnicking, camping and swimming opportunities. Following the lake construction in 1936, black bullheads and largemouth bass were the dominant fisheries. A later attempt to establish trout in the lake was unsuccessful. Currently, bullheads, perch, northern pike, crappie, bluegill and largemouth bass inhabit the lake.

Burke Lake has experienced several partial and two complete winterkills (SD GFP, 1985).

Background/Historical Information

Burke Lake was constructed in 1936 as a Work Projects Administration (WPA) project with trees being planted around the lake in 1949. In 1967, the City of Burke deeded the lake to the State of South Dakota under the understanding that the state would develop the lake as a recreation area.

The City of Burke records indicate that the lake has experienced algae and aquatic vegetation problems since the late 1950's. After receiving complaints expressing concern that the dairy farm located adjacent to the lake was polluting Burke Lake, the DENR inspected the operation in 1986. In March 1986, the DENR requested that a plan for controlling runoff from the dairy feedlot be developed and implemented by the feedlot owner. No action was taken by the owner.



Water Ouality Standards

The water quality standards for Burke Lake are based upon the beneficial uses assigned to the lake by the State of South Dakota. Each beneficial use category has a set of water quality criteria established for it. Burke Lake has the assigned beneficial uses of warmwater semipermanent fish life

propagation, immersion recreation, limited contact recreation and wildlife propagation and stock watering. The water quality criteria for Burke Lake are a summary of the criteria defined for all the use categories using the most stringent values for individual parameters. The water quality criteria for Burke Lake are listed in Table 1.

Table 1. Burke Lake Water	Quality Standards
<u>Parameter</u>	Criterion
Total Chlorine Residual	< 0.02 mg/l
Un-Ionized Ammonia Nitrogen	<u><</u> 0.04 mg/l
Total Cyanide	<u><</u> 0.02 mg/l
Free Cyanide	<u>≤</u> 0.005 mg/l
Dissolved Oxygen	>5.0 mg/l
Undissassociated Hydrogen Sulfide	<u><</u> 0.002 mg/l
pH	>6.5 units and
	< 8.3 units
Total Alkalinity	<u><</u> 750 mg/l
Total Dissolved Solids	<u><</u> 2500 mg/l
Conductivity	≤4000 micromhos/cm
Nitrates	≤50 mg/l (as N)
Suspended Solids	<u><</u> 90 mg/l
Temperature	≤90° F
Polychlorinated Biphenyls	<u>≤</u> 0.000001 mg/l
Fecal Coliform Organisms	≤200 per 100 ml*
숙소화적 생각이 가는 사용을 가는 이 사람들이 가는 것이 가는 기가 사용 것이다. 그는 그는 그 없는	

^{*}Based on the mean of a minimum of 5 samples obtained during separate 24-hour periods for any 30-day period, and this value may not be exceeded in more than 20 percent of the samples examined in the 30-day period. A sample may not exceed 400 per 100 ml in any one sample from May 1 to September 30.

METHODS AND MATERIALS

Lake and Tributary Sampling

Water samples were collected from six sampling sites in the Burke Lake study area (Figure 2). Sites 1-3 were located on tributaries to Burke Lake upstream of the inlets to the lake. Site 4 was located at the outlet of the lake where the lake flows into Coon Creek. Sites 5 and 6 were in-lake sites. The description of sampling site locations is as follows:

- Site 1 Located approximately 50 yards upstream of the confluence of the northern tributary to Burke Lake.

 Latitude: 43° 10' 04"N. Longitude: 99° 15' 34". Site 1 samples are representative of the water quality of the northern tributary inflow to Burke Lake.
- Site 2 Located approximately 50 yards upstream of the confluence of the west tributary to Burke Lake. This site is downstream from a farmyard/feedlot which borders the lake. Latitude: 43° 10' 26"N. Longitude: 99° 15' 44". Site 2 samples are representative of water quality of the west tributary.
- Site 3 Located approximately 50 yards upstream of the confluence of the

southern tributary to Burke Lake. Latitude 43° 10' 50"N. Longitude 99° 15' 32". Site 3 samples are representative of the water quality of the southern tributary.

- Site 4 Located on the east side of Burke Lake at the outlet. Latitude 43° 10 '32"N. Longitude 99° 15' 22". Site 4 samples are representative of the outflow of Burke Lake.
- Site 5 Located approximately 100 yards northeast of sampling site 2 and centrally located in the southwest bay of Burke Lake. Latitude 43° 10' 28"N. Longitude 99° 15' 38". This site provided in-lake data for the southern half of the lake basin.
- Site 6 Located approximately 100 yards west of sampling site 4 in the eastern part of the lake basin. Latitude 43° 10' 34"N. Longitude 99° 15' 28". This site provided inlake data for the northern half of Burke Lake.

The sampling period for tributary and inlake sampling extended from March 28, 1988 through June 26, 1990. During that period, a total of 155 samples were collected from the six sites (Table 2).

SITE #	SAMPLE TIME PERIOD FROM: TO:	# OF SAMPLES
	5/17/89 - 5/17/89	
2	3/28/89 - 6/26/90	27
	3/28/89 - 6/26/90	22
4	3/28/89 - 6/5/90	17
5	3/28/89 - 6/26/90	44
6	3/28/89 - 6/26/90	44
	Total Samples:	155

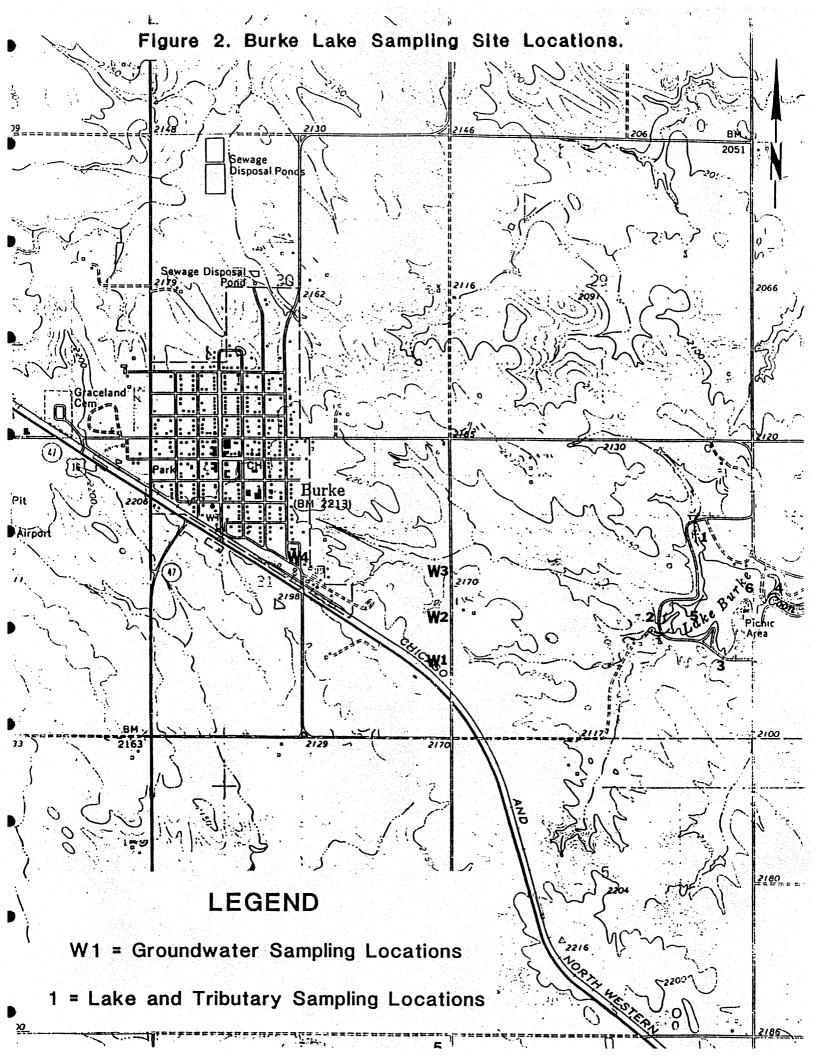


Table 3.-Water Quality Parameters

Water Temperature Total Solids

Air Temperature Total Dissolved Solids

Secchi Disk Total Suspended Solids

Dissolved Oxygen Ammonia

Field pH Nitrates + Nitrites

Fecal Coliform Bacteria Total Kjeldahl Nitrogen

Specific Conductance Total Phosphorus

Laboratory pH Orthophosphate

Total Alkalinity Un-ionized Ammonia

In-lake sites were sampled twice a month at both the surface and bottom. Tributaries and the outflow were also sampled twice a month if flow was present. In addition, tributaries were sampled during rainfall events. Samples were analyzed for the parameters listed in Table 3.

Groundwater Sampling

An abandoned city landfill is located east of Highway 18 outside the extreme southeast limits of the City of Burke. Due to the location of the landfill and geology of the area, groundwater contamination to the lake was suspected. Therefore, a landfill monitoring investigation was initiated.

In early October, 1990, the South Dakota Geological Survey installed three monitoring wells downgradient of the landfill with a hollow auger drilling rig. A City of Burke municipal well was used as an upgradient well (Figure 2). Observation wells were constructed using 2 inch diameter schedule 80 PVC casing and screen. On October 9, 1990, two downgradient wells (W2 and W3) and the

upgradient well (W4) were sampled following a snowfall and melt on October 8, 1990. Also, a sample from the tributary downgradient of the landfill, was collected on the same date for comparison.

Samples were analyzed for a variety of parameters which would indicate groundwater contamination. The following parameters were analyzed: fecal coliform bacteria, conductivity, pH, total dissolved solids, ammonia-N, nitrate + nitrite-N, total Kjeldahl-N, total phosphorus, ortho-phosphorus, and chemical oxygen demand (COD).

Analytical Methods

The laboratory analyses were conducted by the South Dakota State Health Laboratory in Pierre, South Dakota. Lake and tributary sampling was conducted by the South Dakota Game, Fish and Parks Department. Groundwater sampling was conducted by the DENR. Methods used for chemical analyses are listed in Table 4.

Secchi disc* Dissolved oxygen Azide/Winkler pH probe Fotal alkalinity Potentiometric Ammonia-N Phenate Nitrate + Nitrite-N Cadmium reduction Kjeldahl-N Colorimetric Ortho-phosphorus Fotal phosphorus Fotal solids Gravimetric (103-105°C) Fotal dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Parameter	Method	Reference
Dissolved oxygen pH pH probe Total alkalinity Potentiometric Ammonia-N Phenate Nitrate + Nitrite-N Cadmium reduction Kjeldahl-N Colorimetric Ortho-phosphorus Ascorbic acid Total phosphorus Persulfate digestion Total solids Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Temperature	Thermometric	APHA (1985)
pH probe Total alkalinity Potentiometric Ammonia-N Phenate Nitrate + Nitrite-N Cadmium reduction Kjeldahl-N Colorimetric Ortho-phosphorus Ascorbic acid Total phosphorus Persulfate digestion Total solids Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Secchi disc*	Shaded side of boat	Lind (1974)
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Ammonia-N Nitrate + Nitrite-N Cadmium reduction Kjeldahl-N Colorimetric Ortho-phosphorus Ascorbic acid Persulfate digestion Total phosphorus Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	pH	pH probe	APHA (1985)
Nitrate + Nitrite-N Kjeldahl-N Colorimetric Ortho-phosphorus Ascorbic acid Persulfate digestion Total phosphorus Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Total alkalinity	Potentiometric	APHA (1985)
Kjeldahl-N Colorimetric Ortho-phosphorus Ascorbic acid Persulfate digestion Total solids Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Ammonia-N	Phenate	EPA (1983)
Ortho-phosphorus Ascorbic acid Persulfate digestion Total solids Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Nitrate + Nitrite-N	Cadmium reduction	EPA (1983)
Total phosphorus Persulfate digestion Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Gravimetric (103-105°C) Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Kjeldahl-N	Colorimetric	APHA (1985)
Total solids Gravimetric (103-105°C) Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Ortho-phosphorus	Ascorbic acid	EPA (1983)
Total suspended solids Gravimetric (103-105°C) Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Total phosphorus	Persulfate digestion	EPA (1983)
Total dissolved solids Gravimetric (180°C) Fecal coliforms Membraned filter (1989)	Total solids	Gravimetric (103-105°C)	EPA (1983)
Fecal coliforms Membraned filter (1989)	Total suspended solids	Gravimetric (103-105°C)	EPA (1983)
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Conductivity Conductivity probe	Fecal coliforms	Membraned filter (1989)	APHA (1985)
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Chemical oxygen demand Colorimetric	Chemical oxygen demand	Colorimetric	EPA (1983)

Sediment Sampling and Survey

A sediment and water depth survey was conducted on Burke Lake on February 2, 1989. Cross-section transect lines were established on the lake and water and sediment depths were measured with calibrated rebar and recorded at intervals along the transects.

At the same time as the survey, one sediment sample was collected from the center of the lake and submitted to the U.S. Army Corps of Engineers Laboratory in Omaha, Nebraska, for analysis of metals, pesticides and other potentially toxic chemicals.

Biological Sampling

From March, 1989 through March, 1990, water samples were collected on a bi-weekly basis and submitted to Mr. Keith E. Camburn for algal quantifications and identifications.

Watershed Inspection

A watershed inspection was conducted by DENR and U.S. Soil Conservation Service (SCS) personnel in July, 1990. Photographs of potential pollution sources in the watershed were taken in October 1990 by DENR personnel.

RESULTS AND DISCUSSION

In-Lake Water Quality

In-lake water quality mean data are summarized in Table 5 and the complete data set is presented in Appendix A. Discussions of the results have been divided into the components of trophic condition, algae, nutrients, fecal coliform bacteria, and fisheries impacts.

Trophic Condition

Lakes are commonly classified according to trophic condition. Three categories are used: 1) oligotrophic lakes are low in nutrients, are generally clear and support low numbers of animals and plants, 2) mesotrophic lakes are intermediate in nutrient load, and 3) eutrophic lakes are high in nutrients and support large numbers of plants and animals. These lakes often experience excessive algal and weed growth.

Lakes naturally age from oligotrophic to eutrophic conditions over thousands of years. Human activities, however, accelerate the process as a result of increased nutrient and sediment additions.

The Carlson Trophic State Index (TSI) is a method of ranking the trophic condition of the lake based on total phosphorus, Secchi depth and chlorophyll a measurements. The

arameter Comb	ined mean values for sites 5 & 6
Secchi Depth	1.6 ft.
D.O.	11.6 mg/l
Field pH	8.7
Fecal Coliform	11.5 per/100 ml
Conductivity	341
Alkalinity	162 mg/l
Total Solids	274 mg/l
Total Dissolved Solids	238 mg/l
Total Suspended Solids	36 mg/l
Ammonia	0.2 mg/l
Nitrate + Nitrite	0.1 mg/l
TKN	2.6 mg/l
Total PO4-P	0.34 mg/l
Ortho PO4-P	0.07 mg/l

scale runs from 0 to 100 and lakes with values over 50 are considered to be eutrophic. Each division of 10 (10, 20, 30 etc.) indicates a doubling of algal biomass.

TSI's for Burke Lake are based on mean total phosphorus and mean Secchi depth for both sampling sites in the lake over the entire study period. Based upon these values, the TSI value for Burke Lake is 90.8. This value is near the top of Carlson's scale classifying the lake as highly eutrophic.

Algae

Blue-green algae dominated Burke Lake during most of the sampling period. Diatoms were dominant during the cooler periods of spring (Figure 3). This relationship biologically indicates that the lake is highly eutrophic. Of the blue-green algae identified in the lake, the genera Aphanizomenon were the dominant forms present during the study period. Aphanizomenon are commonly identified as problem algae related with eutrophication, taste and odor problems, toxicity and aesthetic nuisance (Taylor, 1974).

Blue-green algae are similar in structure to bacteria but unlike bacteria, are able to photosynthesize like other green plants. Blue-greens often become abundant in nutrient rich lakes because they have several adaptations which allow them to outcompete other algae under enriched conditions. Upon death, blue-green algae cells decompose causing odor and scum formation. Several blue-green algae genera produce toxins which can be poisonous to animals and people (Haynes, 1988).

Other algal groups identified in the study included green algae, cryptomonads, and diatoms. Green algae are a diverse group and are also commonly found in eutrophic lakes. Green algae are generally not reported to produce toxins.

Cryptomonads are a group of motile algae which develop dense populations during cold periods of the year. They seem to do well

under low light conditions and are commonly found in lakes during ice cover conditions.

Diatoms are algae which have silicified cell walls and typically exhibit an early spring population explosion when conditions are too cold for green and blue-green forms. Diatoms can be dominant forms throughout much of the year in eutrophic lakes in cooler regions.

Phosphorus

Phosphorus is an essential nutrient for plant growth and is typically the cause of excessive growth of algae and aquatic plants. Total phosphorus concentrations in excess of 0.03 mg/l have been shown to contribute nuisance growth (NIPC, 1989).

During the entire study period, total phosphorus concentrations exceeded 0.03 mg/l at both in-lake sites (range 0.14 - 1.76 mg/l). Surface and bottom samples exhibited similar phosphorus concentrations indicating the lake remains mixed throughout the year.

These data indicate that the lake has a large phosphorus surplus and therefore is susceptible to algal blooms and nuisance weed growth.

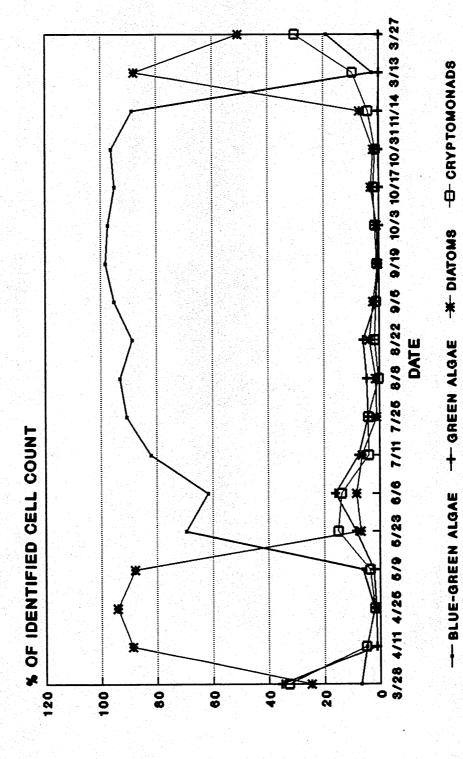
In general, sources of phosphorus to lakes include detergents, human and animal wastes, fertilizers and dead plant materials.

Nitrogen

Nitrogen is often the second most important nutrient to aquatic weed and algal growth in lakes. Inorganic nitrogen concentrations (NO3+NO2+NH4) in excess of 0.3 mg/l are sufficient to stimulate algal growth (Sawyer, 1952).

Inorganic nitrogen concentrations in Burke Lake exceeded 0.3 mg/l throughout much of the 1988 sampling period and occasionally in 1990 (Appendix A). Nitrate + Nitrite (NO3+NO2) values ranged from -0.40 mg/l and ammonia (NH4) values ranged from < 0.02 - 1.48 mg/l throughout the study period.

Figure 3. Algal Cell Counts for Burke Lake



Nitrogen is believed to be the limiting nutrient to algal growth if the ratio of total nitrogen (N) to total phosphorus (P) is less than 10:1 (Shaw, 1989). Blue-green algae are often dominant in lakes with a low N:P ratio since they are able to utilize atmospheric nitrogen which is dissolved in the water. Other algae cannot utilize atmospheric nitrogen and rely on inorganic forms. The N:P ratio in Burke Lake is less than 10:1 thus indicating it is nitrogen limited.

These data suggest that nitrogen concentrations in Burke Lake often exceed that needed to stimulate excessive algal growth, particularly blue-greens.

Major sources of nitrogen to lakes are fertilizers, animal and septic wastes.

Fecal Coliform Bacteria

Fecal coliform bacteria cell counts are used as indicators of human pathogens in water. Although the fecal coliforms are not usually pathogens themselves, they exhibit a close relationship with the human pathogens <u>Salmonella</u> and <u>Shigella</u>.

Recommended microbiological criteria for South Dakota public swimming beaches are that any single sample exceeding 200 fecal coliform/100ml of water is considered as "possibly unsafe" and and two consecutive samples exceeding 300 fecal coliform/100ml is considered unsafe for swimming.

Fecal coliform bacteria concentrations in Burke Lake did not exceed the swimming criteria during the study period. Fecal coliform concentrations ranged from < 10 to 120 per 100ml (Appendix A).

Fisheries Impacts

The beneficial uses assigned to Burke Lake include that of warmwater semipermanent fish life propagation. During the study period, most water quality parameters were sufficient to support a healthy warmwater fishery. However, the warmwater fishery criteria of 0.04 mg/l of un-ionized ammonia was exceeded on

numerous sampling dates in 1989 and on one sampling date in 1990. Un-ionized ammonia concentrations over the study period ranged from 0.00 to 0.15 mg/l (Appendix A).

This data indicates that ammonia concentrations in the lake have reached sufficient concentrations to stress or kill warmwater game fish species. Several fish kills were reported prior to implementing this study. However, no fish kills were reported during the study period.

Tributary Water Quality

During the sampling period, only one sample was collected from tributary site #1 due to the lack of runoff from this tributary. The data collected, however, did not suggest that water quality degradation was occurring in this tributary.

Tributary sites #2 and #3 were sampled a total of 27 and 22 times respectively (Table 2). Water quality data indicated that these tributaries experience high levels of contamination and in return deliver high concentrations of contamination to Burke Lake.

Water quality parameters which are most closely associated with degrading the recreational potential of the lake and are also most indicative of pollution sources include fecal coliform bacteria and nutrients.

As previously discussed, fecal coliform concentrations exceeding 200/100ml of water in lakes is considered potentially unsafe. Although this standard does not apply directly to streams, it does help display the high level of contamination occurring in the tributaries. During the study period, fecal coliform concentrations in tributary sites #2 and #3 frequently exceeded 200/100ml. The range of fecal concentrations for sites #2 and #3 ranged from < 10 to 4,000,000 and from < 10 to 6,300, respectively (Appendix A).

Also, as previously discussed, excessive nutrient concentrations are the primary cause of excessive blue-green algae and aquatic plant growth in lakes. Tributaries #2 and #3 both contributed high concentrations of

phosphorus and nitrogen to Burke Lake consistently throughout the study period. Total phosphorus concentrations for the tributaries ranged from 0.21 - 18.90 mg/l at site #2 and from 0.08 - 2.56 mg/l at site #3. Inorganic nitrogen concentrations ranged from 0.02 - 2.80 mg/l at site #2 and from 0.02 to 15.24 mg/l at site #3 (Appendix A).

Other water quality problems found in the tributaries also potentially have negative impacts on the fish resources in Burke Lake. Some of the problems include low dissolved oxygen concentrations, high un-ionized ammonia, and high concentrations of suspended solids.

Groundwater Investigation

The landfill property consists of 6.66 acres of which 4.09 acres were purchased in 1928 and the balance was purchased in 1967. The landfill was used for municipal wastes until sometime in late 1974 or early 1975.

The surficial geology in the area of the landfill includes deposits of the Herrick, Ash

Hollow and Valentine Formations. These formations consist of coarse, fine and very fine sands. The Ogallala aquifer lies within the Valentine formation which is composed of very fine to fine sand with some silt and clay lenses (Hammond, 1990).

The landfill is located approximately 5000 feet upgradient of the lake and approximately 1000 feet upgradient of a lake tributary. The direction of groundwater flow from the landfill is to the east and northeast moving towards the lake (Hammond, 1990).

The results of the October, 1990 groundwater and tributary sampling are summarized in Table 6. The data indicate that groundwater in the watershed has elevated concentrations of nitrogen and phosphorus. A previous study (Hammond, 1990) reports that all city wells located within city limits display high concentrations of nitratenitrogen. This study confirmed the elevated nitrate in a city well but also indicated elevated nitrate and phosphorus in the monitoring well downgradient of the city and landfill.

Parameter	Background Well	MW #2	MW #3	Trib.
Fecal Coliform	14	<2	2	3800
pH	7.57	7.22	7.91	7.52
Cond.	400	588	382	515
TDS	288	395	288	363
Ammonia	< 0.02	0.03	0.03	0.39
NO3+NO2	6.1	2.5	7.9	< 0.1
TKN	0.14	0.58	0.88	1.56
Total PO4-P	0.047	0.082	0.359	0.966
Ortho PO4-P	0.030	0.073	0.350	0.367
COD	22	30	30	40

It is not clear what the exact source or sources of the contamination are from this investigation, however, data from this study and from the Hammond investigation indicate that the source(s) of the groundwater contamination is likely located within or near the southeastern city limits.

Although nutrient levels in the groundwater contribute to problems in the lake, the nutrient levels found in groundwater were not substantial enough to account for the gross nutrient contamination of the tributaries. In addition, fecal coliform and ammonia contamination in the tributaries is not attributable to groundwater. Therefore, other watershed sources are the primary sources of pollution to the tributary and to the lake.

Sediment Survey

The DENR conducted a sediment and water depth survey of Burke Lake on February 16, 1989. Cross-section locations used for the survey and the respective cross-section graphs are presented in Appendix B.

The maximum water depth found during the survey was 14.2 feet in the northcentral part of the lake (Figure 4). Most nearshore areas (50 feet from shore) have water depths less than 4 feet deep.

The mean sediment depth in the lake is 2.76 feet with a maximum sediment depth of 7.2 feet occurring in the central portion of the lake (Figure 5). Most nearshore areas have sediment depths of less than 2 feet deep. The lake has as an approximate sediment volume of 125,000 cubic yards.

High concentrations of total suspended solids were recorded entering the lake from tributary sampling site #2 (Appendix A) and typically corresponded with elevated organic nitrogen values (TKN - Ammonia) indicating that the solids are at least partially organic. It is unclear to what extent the load is made up of inorganic sediment (i.e. soil particles). Due to the ease at which sediments likely become redistributed in such a small shallow

lake, it is unclear which tributary is the major source of sediment loading to the lake.

Elutriate Analysis

A sediment sample collected from the center of the lake was analyzed for a variety of metals, pesticides and other potentially toxic chemicals. Results indicated no toxic levels of chemicals in the elutriate (Appendix C). Therefore, the analysis indicated that disturbance of sediments via dredge activities should not present a toxicity problem to the lake.

Watershed Inspection

On July 31, 1990, DENR and SCS representatives conducted a watershed survey to identify potential sources of pollution to Burke Lake. Several potential sources were identified including a dairy operation, a horse pasture, a mule pasture, a fertilizer storage facility, and an abandoned landfill.

Photographs documenting watershed pollutant sources were taken on October 3, 1990 and are included in Appendix D.

Dam Inspection

The dam was inspected on May 8, 1990 by the DENR Division of Water Rights. The DENR recommended that the following repair and maintenance activities be performed:

- 1. Trees, brush and the root systems of the larger trees should be removed from both slopes of the dam and the holes backfilled with impervious material.
- 2. Erosion of the upstream slope should be repaired and sufficient riprap should be placed along the slope to prevent further erosion.
- 3. Trees and brush should be removed from the spillway discharge channel and the channel should be leveled to provide uniform flow.
- 4. Major spalls, cracks, and deteriorated joints should be repaired.

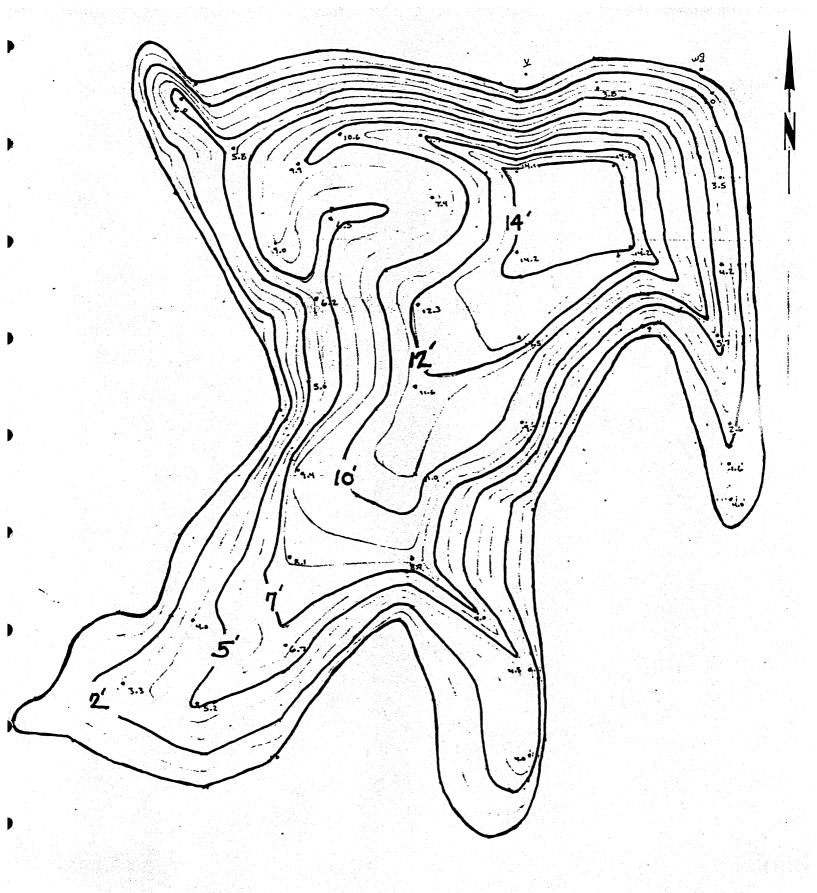


Figure 4. Burke Lake Bathymetric Map.

Note: Survey Data Collected 2/16/89. Scale: 1 INCH = 180 FEET

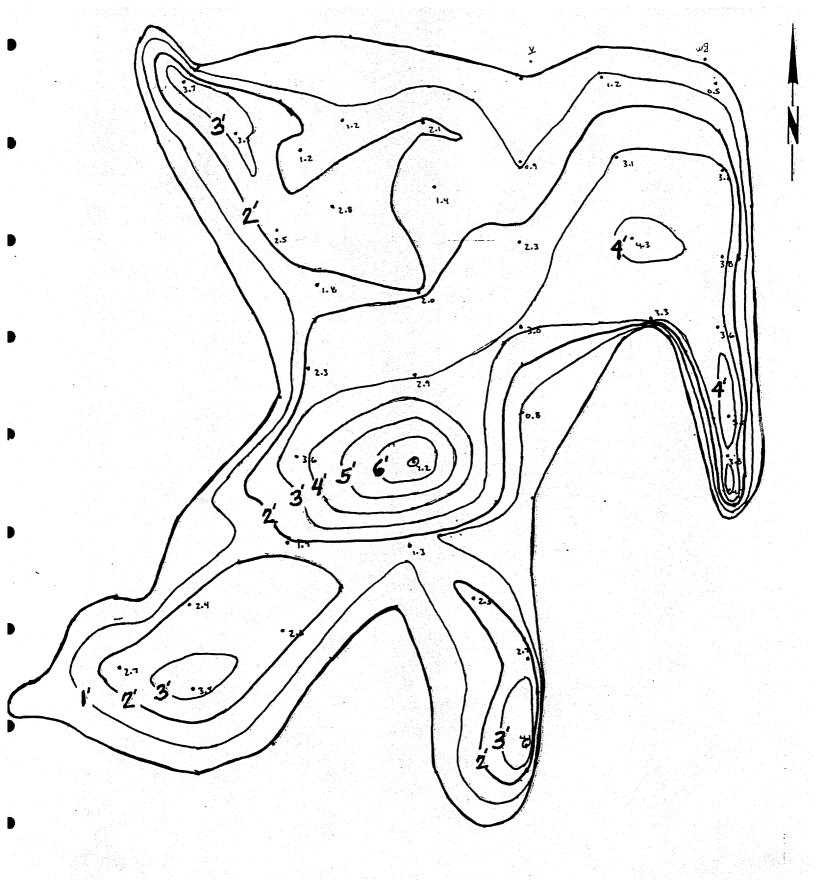


Figure 5. Burke Lake Sediment Thickness Contour Map.

Note: Survey Data Collected 2/16/89. Scale: 1 INCH = 180 FEET

CONCLUSIONS

Burke Lake is highly eutrophic due to excessive nutrients entering the lake. Nutrients are entering the lake primarily from the west and south tributaries. Nutrients carried within the general groundwater flow to the lake are also likely contributing to the nutrient enrichment of the lake. However, the nutrient contribution from groundwater is likely much less than that from the west and south tributaries.

Potential sources of nutrients to all the tributaries included a dairy operation, a horse pasture, a mule pasture, and cropland. Fecal coliform data suggests that animals are the primary source of contamination in the west and south tributaries. These data in conjunction with the findings of the watershed inspection indicate that the dairy operation, located adjacent to the west and south tributaries, is the primary source of nutrient and fecal contamination entering the west and south tributaries. A horse pasture, located adjacent to the west tributary, is an additional source of nutrient and fecal contamination in the west tributary.

Due to the limited flow of the north tributary during the study period, it is not clear if this tributary could contribute significant contamination to the lake in times of heavy precipitation. However, the watershed inspection suggests that a mule pasture located along the north tributary could potentially contribute significant contamination to the tributary due to the improper storage of manure piles on tributary slopes.

The source or sources of groundwater nutrient contamination in the western portion of the watershed appear to be located within or near the southeast city limits. Potential sources of the contamination included a fertilizer storage facility and the abandoned landfill. A more comprehensive groundwater monitoring study would be needed to thoroughly evaluate the sources of groundwater contamination and the impacts of groundwater quality on the lake.

Based upon the evidence from this study, the most threatening source of pollution to the lake is the dairy operation.

Data from the sediment survey and from water quality sampling suggest that sediment loading is not a major problem in Burke Lake. However, due to the small and shallow character of the lake, the nutrient rich sediments likely have a large impact on algal and aquatic weed growth in the lake and also likely contribute to winter fish kills due to lack of sufficient depth.

RESTORATION ALTERNATIVES

Watershed Restoration

All restoration alternatives should begin with addressing contamination sources within the watershed. Reducing pollutant loading from the watershed will accelerate the control of aquatic vegetation in the lake, improve the fishery through the reduction of winterkill, and will improve swimmability. Watershed restoration activities should include those discussed below.

Foremost, an adequate system to control pollution coming from the dairy operation needs to be developed and implemented. In addition, pasture and cropping activities within the watershed need to be addressed via information to land owners on how to minimize manure, fertilizer and sediment contributions to water resources. Finally, sources of nutrient contributions to groundwater in the watershed need to be identified and controlled. A more comprehensive groundwater study may be needed to identify sources.

Although sediment loading to the lake does not appear to be as accelerated as in many South Dakota lakes, it is a problem which threatens the fishery and contributes to the eutrophic condition of Burke Lake. Addressing watershed activities to control sedimentation via education of land owners is the first step. Should this prove ineffective, a sediment/nutrient control system may have to be considered. An example of such a system would be the installation of one or more sediment detention basins or constructed wetlands within the watershed to act as sediment and nutrient traps and filters.

Prior to the implementation of any inlake restoration activities, the lake's dam should be renovated and approved by the DENR Division of Water Rights. Renovation recommendations were outlined previously.

In addition to the Water Rights recommendations, cattails, sediment and debris should be removed from the area just in front of the upstream dam lip.

In-Lake Restoration

In-lake restoration can be approached either actively or passively. The passive approach is to take no action in the lake. If watershed contributions of nutrients to the lake are controlled, the lake may slowly improve in water quality over a period of several years. How much the lake would improve and to what extent it would improve are unknown. However, it is likely the lake would continue to experience algal and aquatic vegetation problems for many years.

Active in-lake restoration alternatives will be discussed in regards to the desired goals. The goals discussed are algae/weed control, fishery improvement and swimmability improvement.

Algae/Weed Control

Algae/aquatic weed control is accomplished by reducing the amount of nutrients available for plant growth or by direct removal of plants.

Sediment Nutrient Reduction

Nutrient availability from lake sediments can be reduced by a variety of methods including sediment consolidation, bottom sealing, sediment oxidation and full or partial lake dredging.

Sediment consolidation involves draining a lake and allowing the bottom sediments to dry out and become hardened. As a result, nutrients become bound in the hard sediment and sediment volume decreases due to drying. Sediments may take from several months to more than a year to dry sufficiently. Hardening the sediments may decrease or completely eliminate groundwater recharge to a lake. Since groundwater is an important source of water to Burke Lake, this method is not recommended.

Bottom sealing involves placing plastic or other sealing materials over the sediment surface to stop the exchange of nutrients between the sediments and water. Oxidation involves treating the upper sediment layer chemically to enhance the removal of nutrients. Of the methods described, only dredging has been demonstrated to be an effective long term method of sediment nutrient control.

Dredging involves the physical removal of nutrient rich sediments from the lake bottom. Full lake dredging involves the removal of sediments from most or all of the lake. Selective dredging involves the removal of sediments from only specific areas of the lake. For example, sediments could be removed from specific shorelines in order to create improved depth for swimming and to control aquatic vegetation growth.

Selective dredging will aid in the control of growth of aquatic weeds and algae, increase lake accessibility to boaters and increase swimmability. This approach will likely be less costly than full lake dredging but will also likely not result in as significant a reduction of algae and weeds. As with full lake dredging, care must be taken to maintain adequate shallow water habitats for fish production.

Dredging can be accomplished by hydraulic dredging, drag-line dredging, or mechanical dredging with a backhoe following a partial or complete drawdown of the lake. In general, drag-line and mechanical dredging are most economical if the volume of sediment to be removed is small. These methods of dredging are more labor intensive than hydraulic dredging and involve trucking the removed sediments away to a disposal area. In addition, these methods are more time intensive than hydraulic dredging.

Average costs to conduct mechanical dredging on two lake projects in South Dakota were \$2.63 per cubic yard of sediment. Average costs for hydraulic dredging by the State of South Dakota are \$1.46 per cubic yard with a 10 inch dredge.

Hydraulic dredging is most economical when large volumes of sediment are to be removed because of the initial costs involved in transporting the dredge, setting up the dredge, and constructing sediment holding ponds. Sediments mixed with water are pumped through a pipe to a sediment disposal pond.

The State of South Dakota currently owns and operates four dredges. By doing so, the state has been able to substantially reduce the cost of hydraulic dredging in comparison to contracting with a private dredging company. The cost for the state to operate a 8 inch dredge for one year is approximately \$220,000. This cost includes all labor and disposal pond construction costs. Private contracting costs per cubic yard of sediment are approximately two to three times higher than the state cost.

Water Nutrient Reduction

Nutrient availability from the lake water itself can also be reduced. The best demonstrated method for accomplishing this is alum (AlSO4) treatment. Alum can be added to the water under controlled chemical conditions. The alum reacts with phosphorus in the water and forms an insoluble precipitate that settles into the sediments. This method works best in hard water lakes that are thermally stratified but it has been demonstrated to be effective for shorter durations in shallow mixed lakes. The alum must be applied in highly controlled conditions in order to prevent formation of substances that are harmful to aquatic life. Also, application of enough alum to coat the lake bottom with a thick layer of alum is desirable for sediment nutrient inactivation and longer duration water nutrient inactivation.

Alum treatment is small shallow lakes can result in the presence of the precipitate or floc on shores and on the lake bottom. This may be undesirable to lake users. In addition, alum treatment can result in a increase in water clarity which in turn can encourage an initial increase in aquatic weed growth.

The major cost involved in alum treatment is generally labor. A 40 acre lake in Ohio was treated with 100 tons of alum at a cost of \$14,000 for the alum alone. The treatment costs will decrease if lower doses of

alum are used as would be the case if the objective is to remove phosphorus from the water column only.

Aguatic Plant Removal

Aquatic plants can be removed from the lake through harvesting or chemical treatments. Chemical treatments typically result in short term aesthetic results and do not aid in the removal of nutrients from the system. In addition, toxic chemicals used in the treatments can accumulate in the lake sediments.

Harvesting removes weeds from the lake and as a result removes nutrients from the system. Harvesting costs range from approximately \$140 to \$310 per acre.

Harvesting can be used alone or in addition to dredging. However, if a dredge program is properly planned, the dredge activity can remove nuisance vegetation and create sufficient water depth to limit further aquatic weed growth from selected areas.

Fishery Improvement

Fishery improvement can be accomplished through preservation and enhancement of critical habitats (i.e. nursery, food production and winter habitats) and through the attainment of a balanced fishery (predator/prey balance). Nursery and food production habitats can be identified, preserved and enhanced and a balanced fishery can be attained via cooperation with and recommendations from the South Dakota Game Fish and Parks Department.

Critical habitat can be enhanced to reduce the risk of winterkills by creating sufficient deep water areas in the lake and by removing excessive aquatic vegetation and organic sediment materials. This can be accomplished via dredging.

Swimmability Improvement

Swimmability of the lake can be improved through control of nuisance vegetation. Methods for the control of vegetation were discussed previously.

RECOMMENDATIONS

Recommendations for the restoration of Burke Lake will be listed in order of priority, where the highest priority recommendations are listed first and should be implemented prior to, or in conjunction with, the implementation of lower priority recommendations.

1. An adequate system to control pollution coming from the dairy operation needs to be developed and implemented. The land owner working in cooperation with the Soil Conservation Service, should be responsible for the installation of adequate pollution control structures. The operation must achieve compliance with ARSD 74:03:18 regulations pertaining to animal feeding operations.

The design and installation of two animal waste control ponds was estimated by SCS to cost approximately \$150,000.

- Pasture and cropping activities within the watershed need to be addressed via provision of information to land owners on how to minimize manure, fertilizer, and sediment contributions to water resources.
- 3. Sources of nutrient contributions to groundwater in the watershed should be identified and controlled. As a start, the South Dakota Department of Agriculture should be notified to conduct a compliance inspection at the grain elevator/fertilizer storage facility located near the southeast city limits. In addition, a more comprehensive groundwater study is needed to identify sources.

Minimally, it is recommended that the wells sampled in the D\F study be resampled in the spring after at least one moderate spring rainfall event. In addition, two additional wells should be installed and sampled at the same time as the others. One well should be installed upgradient of the landfill and downgradient of the grain elevator/fertilizer storage facility in order to distinguish potential nutrient impacts from the

elevator facility. A second well should be installed downgradient of the landfill and upgradient of the adjacent pasture. This well will serve to distinguish between impacts from the landfill and those from the pasture. Another upgradient well which can be sampled alike to the other downgradient wells, should also be sampled.

Installation of two new wells by SDGS and one time sampling of four wells by the DENR would cost approximately \$2,000.

- 4. Prior to the implementation of any in-lake restoration activities, the lake's dam must be renovated by undergoing the following repairs and maintenance:
 - Trees, brush and the root systems of the larger trees should be removed from both slopes of the dam and the holes backfilled with impervious material.
 - Erosion of the upstream slope should be repaired and sufficient riprap should be placed along the slope to prevent further erosion.
 - Trees and brush should be removed from the spillway discharge channel and the channel should be leveled to provide uniform flow.
 - Major spalls, cracks and deteriorated joints should be repaired.
 - Cattails, sediment and debris should be removed from the area just in front of the upstream dam lip.

GFP working in conjunction with the local sponsor should be responsible for dam repairs.

- 5. In-lake restoration is recommended to include the following:
 - Full lake dredging to remove sediment nutrients and nuisance stands of aquatic vegetation, and to create

deep water areas which will limit regrowth of aquatic vegetation and create deep fish habitat holes to prevent winterkills. Since lake sediments contain and often recycle nutrients to the lake, they represent a pollution source to the lake. Sediment removal via dredging is often utilized as a way of improving the lake quality through removal of nutrients and enhancement of lake usability by increasing depth. Lake deepening can increase storage capacity, prevent winterkills, expand recreational areas, decrease area available for plant growth and may aid in the reduction of resuspended sediments (Sefton, 1983). A fish management plan should be developed with the GFP department prior to any dredging activities.

Full lake dredging by DENR is estimated to take approximately 1.5 years to complete with an 8 inch dredge. Dredging costs are estimated to be approximately \$330,000.

- Following dredging activities, a one time alum treatment would aid in reducing nutrient availability from the lake water itself. This method works best in hard water lakes that are thermally stratified. After the completion of dredging and the creation of deep holes, the lake will likely stratify to some extent and will therefore likely benefit from alum. The alum must be applied in highly controlled condi-

tions in order to prevent formation of substances that are harmful to aquatic life. Also, application of enough alum to coat the lake bottom with a thick layer of alum is desirable for sediment nutrient inactivation and longer duration water nutrient inactivation.

The cost of alum and application by a private contractor is estimated to be approximately \$12,000. Additionally, the cost of contractor travel is estimated to be up to \$10,000 based upon an applicator located in Pennsylvania. Travel costs per lake can be substantially reduced if more than one lake is treated during the trip.

- 6. Upon completion of the outlined lake restoration techniques, an ongoing lake management/protection plan for the lake should be developed and implemented. A system of long term water quality monitoring should be developed as part of that strategy to detect problems early. This plan should be developed as a cooperative effort between GFP, DENR, and the City of Burke.
- 7. If aquatic vegetation continues to persist at nuisance levels after dredging is completed, selective weed harvesting may be used as an option in the long term lake management strategy. A small scale harvester such as a Hockney could be purchased through cooperative funding and operated by a local or state entity. The purchase price of a Hockney is approximately \$10,000 without shipping.

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APPENDIX A
Water Quality Data

	UNIAM	0.003	0.005	0.000	0.000	0.005	0.000	0.001	0.000	0.000	0.002	0.412	900.0	0.097	0.004	0.00	0.042	0.005	900.0	0.002	0.001	0.000	0.000	0.00	0.000	0.00	0.000	0.021	0.001	0.005		0.002					0.003	0.000	0000	0000	
	0P04-P	0.02	0.17	0.15	0.18	0.03	0.42	0.37	0.20	0.29	0.64	0.19	0.75	0.13	1.02	1.17	1.95	0.87	0.71	0.71	0.45	90.0	0.05	0.07	0.14	0.12	0.18	0.14	0.21	0.19		0.17					0.54	0.28	0.22	0.17	
	TP04-P	0.10	0.25	0.41	0.21	1.58	0.76	2.16	0.85	0.39	1.17	89.9	1.54	18.90	1.63	1.83	5.32	1.07	0.85	0.88	0.61	0.13	0.10	0.08	0.19	0.17	0.32	0.24	2.04	0.33		0.34					0.91	0.28	0.25	0.31	
	TKN-N	0.82	1.38	1.78	0.81	3.94	1.68	1.78	2.14	1.08	3.46	6.76	1.8	27.20	2.10	3.03	9.59	1.22	1.00	1:1	1.20	0.58	0.63	0.35	0.64	0.52	1.18	1.77	2.19	1.27		1.49					2.81	09.0	69.0	0.52	
11	N.		1.20																													<0.10					c0.10	c0.10	<0.10	c0.10	
	2		0.47																		1,74		54.00									0.00					- 2		c0.02		
			16												1.																	53					21	9	12	97	
	TDSOL	401	315	317	314	210	319	334	355	391	407	417	398	510	398	430	535	390	387	388	340	241	263	546	307	307	357	358	321	261		437					887	357	371	319	
	TSOL	405	331	346	349	251	371	503	1048	423	455	5799	729	1863	797	526	2187	392	397	401	395	526	268	283	336	311	807	329	634	283		994					209	363	383	365	
	ALK(T)	115	236	230	219	158	232	240	225	293	197	351	255	491	567	588	306	257	242	240	234	185	194	191	241	231	529	271	220	17		281					277	232	238	542	
	LABPH	8.37	7.77	7.70	7.82	8.09	8.03	7.31	7.42	7.89	7.50	7.15	7.36	29.9	7.29	7.38	7.22	7.52	7.85	79.7	7.82	2.96	7.66	7.98	7.77	7.78	7.56	7.78	7.47	7.70		8.00					7.63	7.73	7.61	7.84	
	COND																																								
	FECAL	9	₽	2	07	2100	1600	20	8	180	720000	4000	1000	2000	4000	7100	4000000		790	1300	15000	0₽ ₽	0 ₽	€	70	30	1700	1200	30	1600		10					6300	009	~10	ន	
	표	8.15	7.70	9.60	7.10	7.40	7.60	7.30	6.70	8.50	7.60	8.60	2.40	8.10	7.10	7.40	8.10	7.90	8.40	7.90	7.80	7.80	09.9	7.40	6.90	7.30	7.10	9.80	7.30	8.70		7.80					8.00	8.10	8.00	8.20	
	DISOX	7.6	8.5	9.0	9.6	5.5	13.5	4.3	2.5	6.3	5.0					2.8	4.5		7.0	7.0	8.5	12.0	7.6	14.0	5.3	10.5	3.5	6.0	4.3	2.0		4.5						7.5	8.0	10.0	
	ATEMP SDISK																																								
	ATEMP	3	9	70	7,6	9	62	R	82	85	29	8	8	3	92	22	8	77	0,7	20	34	9	38	45	25	63	65	22	82	3		2					42	07	95	38	
	WTEMP	8	54	36	37	58	26	8	22	82	99	*	88	8	3	39	65	95	38	36	33	52	38	36	28	26	63	88	8	62		3	ш	ш	ш	ш	43	36	36	33	
	DEPTH	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5		TO SAMPLE	0.5	TO SAMPLE	TO SAMPLE	TO SAMPLE	TO SAMPLE	0.5	0.5	0.5	0.5	
	SAMPLE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE			SURFACE	NO WATER 1	NO WATER 1	NO WATER 1	NO WATER 1	SURFACE	SURFACE	SURFACE	SURFACE																						
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	DATE	05/17/89	03/28/89	03/30/89	04/11/89	04/52/89	05/09/89	05/23/89	68/90/90	06/20/89	06/26/89	07/11/89	07/25/89	08/08/89	08/22/89	09/02/89	09/19/89	10/03/89	10/17/89	10/31/89	11/14/89	03/28/89	03/30/89	04/11/89	04/52/89	05/09/89	05/23/89	68/90/90	06/20/89	06/26/89	07/11/89	07/25/89	08/08/89	08/22/89	09/02/89	09/11/89	10/03/89	10/17/89	10/31/89	11/14/89	

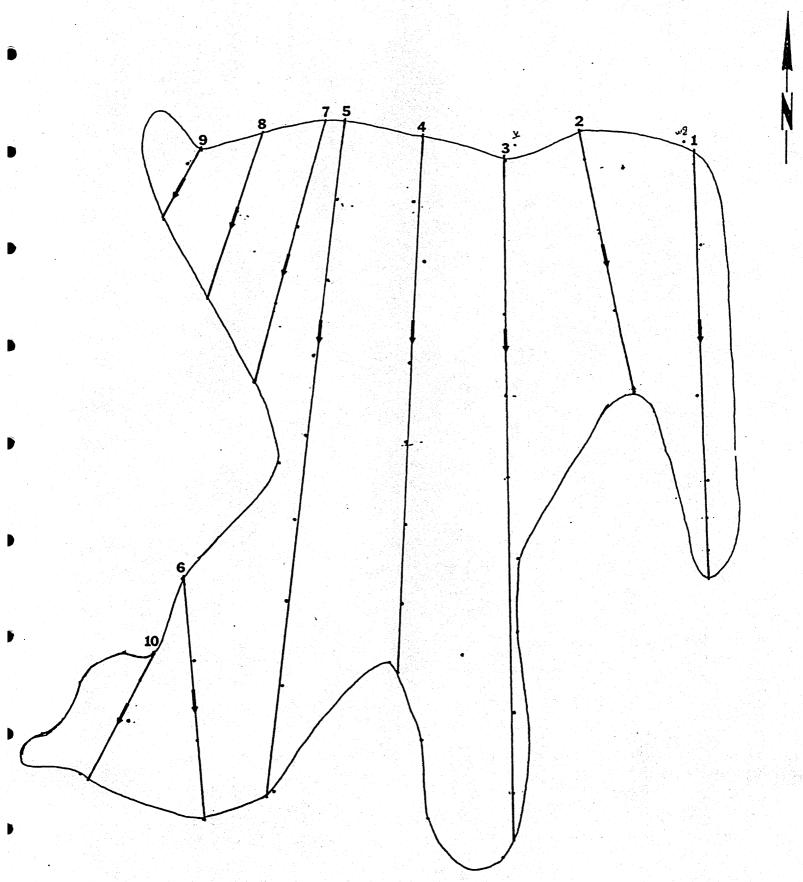
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UNTAM	0.001	0.000	0.003	0.001	0.001	0.007	0.707		0 014	820	3									0	0000	0.00	0.00	0.00	0.00	0.00	0.001	0.002	0.002	0.018	0.030	0.151	0.026	0.046	0.029	0.063	0.042	0.014	0.019	0.013	0.013	0.014	0.015	
d-70d0	0.06	0.03	0.01	0.70	0.02	0.02	0.21	0.19	50.0	60.0	3									90.0	0.05	0.01	0.01	0.01	0.01	0.01	*0.0	90.0	0.05	0.01	0.01	0.02	90.0	0.10	0.10	0.11	0.12	0.10	0.08	0.15	0.13	0.11	0.13	
TP04-P	0.30	0.31	0.53	1.14	0.21	1.38	2.46	2.34	2	7 6	}									72.0	0.25	0.17	0.16	0.17	0.19	0.16	0.18	0.24	0.24	0.16	0.16	0.31	0.36	0.50	0.48	0.48	0.48	0.52	0.51	0.55	0.55	0.52	0.56	
TKN-N	1.76	2.12	2.54	1.83	1.28	5.05	5.24	5.56	2.00	22.2										2.02	1.72	1.48	1.43	1.66	1.45	1.52	1.55	1.77	1.46	5.09	2.52	2.32	2.58	2.96	2.74	3.95	4.54	4.89	4.56	4.76	76.7	5.34	4.67	
N03+2	<0.10	0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	60.10										<0.10	0.10	<0.10	<0.10	6.10	60.10	60.10	<0.10	<0.10	<0.10	<0.10	*0.10	<0.10	<0.10	<0.10	<0.10	.0.10	0.10	0.30	<0.10	<0.10	<0.10	<0.10	<0.10	
AMMON	0.04	0.03	0.04	0.05	0.03	0.15	0.71	1.25	0.04	0.03	}									0.04	0.02	<0.02	<0.02	<0.02	<0.02	<0.02	0.02	0.02	0.02	0.10	0.14	97.0	90.0	0.08	0.10	67.0	0.28	0.05	90.0	20.0	0.11	0.03	0.03	
TSSOL	82	20	3	115	2	1305	1544	576	22	1										18	75	??	51	36	26	91	32	17	31	4	12	12	£	=	2	25	45	7	К	83	53	25	87	
TDSOL	236	546	234	323	220	202	227	201	186	208										237	544	240	237	202	202	227	554	526	221	190	179	506	500	218	215	506	506	206	202	222	522	245	242	
		569				-	-													255	258	291	288	243	564	243	256	267	252	194	191	218	234	553	585	253	251	277	280	285	278	297	329	
ALK(T)	169	186	171	242	3	132	151	148	123	132										171	178	13	177	79	36	172	171	156	154	125	126	130	131	133	134	140	139	143	143	146	148	152	155	
LABPH	8.11	8.19	8.86	7.75	8.40	7.37	7.51	7.32	8.06	9.21										8.21	8.13	8.87	8.88	8.66	8.72	8,48	8.37	8.29	8.57	9.11	20.6	8.84	8.72	8.67	8.69	7.95	8.56	9.11	9.14	8.85	8.68	9.25	9.23	
COND										270										363	363	355	355	330	330	354	356	330	331	526	529	281	270	585	567	306	306	303	303	308	339	307	307	
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.	8.00	7.30	8.70	7.90	8.10	8.10	11.80		9.20	10.30										7.70	7.70	8.90	8.80	8.70	8.70	8.20	8.20	8.60	8.60	8.70	8.80	8.90	9.10	07.6	8.90	8.50	8.70	8.90	9.00	8.70	8.50	9.30	07.6	
DISOX	17.5	11.3	14.2	6.8	11.5	5.5	2.0		2.5	8.0										14.0	17.0	15.6		9.3	9.0	10.5	9.5	14.8	10.5	12.5	12.0	7.5	4.5	5.0	6.0	8.0	7.0			8.0	6.0	15.0	14.0	
ATEMP SDISK										1.3										2.3	2.3	2.3	2.3	5.0		5.0	2.0	2.0	2.0	5.0	2.0	1.3	 	1.0	1.0	0.8	0.8	0.5	0.5	0.5	0.5	6.0	6.0	
ATEMP	89	38	45	09	62	2	22		3	80										58	58	77	7,7	9	9	52	25	3	75	22	22	87	87	8	80	2	2	22	22	22	22	92	92	
VTEMP	54	7,7	43	9	62	88	92		65	78	ш	ш	ш	ш	ш	ш	w	111	101	45	45	7,7	7,5	3	62	9	28	63	63	2	2	2	2	2	7,4	22	65	22	٦	Z	2	2	69	
DEPTH	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	0.5	TO SAMPLE	TO SAMPLE	TO SAMPLE	TO SAMPLE	TO SAMPLE	TO SAMPLE	TO SAMPLE	O SAMPLE	TO SAMPLE	0.5	4.5	3.0	0.5	0.5	7.0	0.5	4.0	0.5	4.0	0.5	4.0	4.0	0.5	7.0	0.5	3.0	0.5	3.0	0.5	3.0	0.5 7.	0.5	2.0	
SAMPLE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	SURFACE	NO WATER T	WATER	WATER	WATER	WATER	WATER	NO WATER T	NO WATER TO	NO WATER T	SURFACE	BOTTOM	BOTTOM	SURFACE	SURFACE	BOTTOM	SURFACE	BOTTOM	SURFACE	BOTTOM.	SURFACE	BOTTOM	BOTTOM	SURFACE	BOTTOM	SURFACE	BOTTOM	SURFACE	BOTTOM	SURFACE	BOTTOM	SURFACE	SURFACE	BOTTOM	
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DATE	03/28/89	03/30/89	04/11/89	04/52/89	05/09/89	05/23/89	06/06/89	06/20/89	06/26/89	07/11/89	02/25/89	08/08/80	08/22/89	09/02/89	09/19/89	10/03/89	10/17/89	10/31/89	11/14/89				04/11/89				05/09/89	05/23/89	05/23/89			07/11/89	Α.										09/19/89 1	

0.048 0.058 0.072 0.072 0.072 0.075 0.075 0.075 0.000 6.01 60.01 6.00 10.00 10.00 0.13 0.14 0.18 0.17 0.20 0.20 0.20 0.06 00.29 00.29 00.52 00.53 00.53 00.53 00.53 00.53 00.64 00.64 3.70 3.71 3.75 4.04 2.98 2.87 3.06 2.93 1.80 ... 6.6.6 6.5 6.5 6.0 **60.10 60.10** 6.10 0.10 **60.10** <0.10 <0.10 6.0 6.10 **60.10** 6.19 0.10 **60.0** 6.10 6.10 **60.10** 0.35 0.08 0.04 0.04 0.05 0.05 0.05 0.02 0.38 0.38 60.02 0.03 0.02 0.03 0.04 0.12 0.05 <0.02 <0.02<0.02 0.03 662266 9.20 7.60 9.20 9.20 9.20 8.80 8.80 8.20 8.20 9.00 9.00 9.00 9.00 8.70 16.0 11.5 12.8 12.5 15.0 17.5 14.8 11.5 11.5 8.0 BOTTOM SURFACE BOTTOM BOTTOM SURFACE BOTTOM BOTTOM BOTTOM BOTTOM SURFACE SURFACE SURFACE BOTTOM SURFACE SURFACE SURFACE BOTTOM SURFACE SURFACE BOTTOM SURFACE BOTTOM BOTTOM SURFACE BOTTOM SURFACE BOTTOM 08/22/89 10/17/89 06/06/89 07/11/89 07/11/89 0/03/89 10/03/89 10/17/89 10/17/89 10/31/89 05/23/89 0/31/89 11/14/89 04/11/89 04/52/89 9/05/89 9/02/89 19/19/89 09/11/80 0/03/89 DATE 10/03/89 11/14/89 04/11/89 04/22/89 05/09/89 05/09/89 05/23/89 68/90/90 07/25/89 07/25/89 08/88/89 08/80/80 08/25/80 03/28/89 3/28/89

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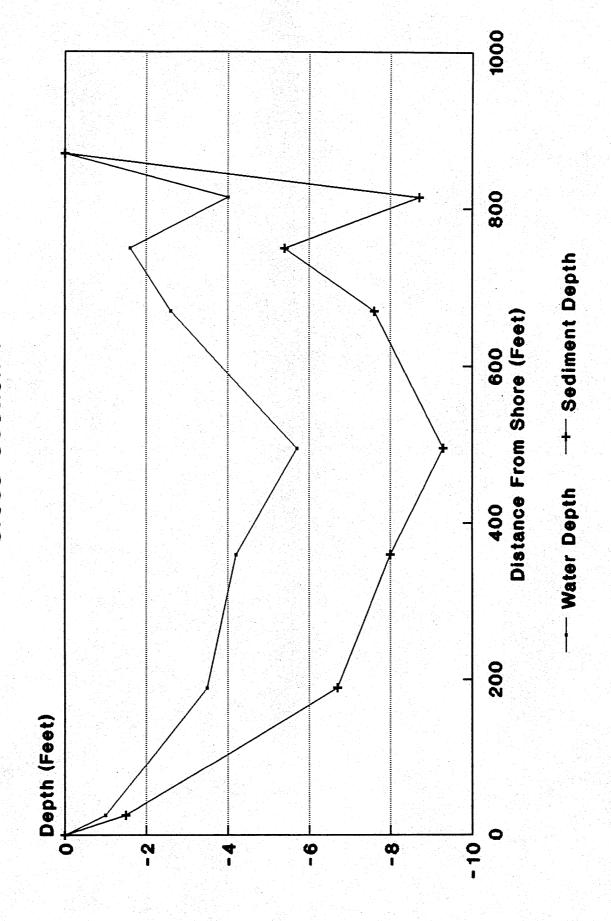
UNIAM	0.001	0.00	0000	0.00	0.000	0.000	0.000	0000	0.00	0000	0.004	0.000
0P04-P	<0.01	<0.01	<0.01	*0.0	0.01	0.02	0.04	0.05	0.02	0.03	0.23	0.13
TP04-P	0.31	0.22	0.18	0.23	0.21	0.16	0.21	0.17	0.21	0.20	0.41	0.37
TKN-N	2.07	2.27	1.71	1.60	1.54	1.48	1.40	1.40	1.37	1.32	2.11	1.8
N03+2	60.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10	<0.10
			0.02									
	1500		75									
			262									
TSOL	295	296	278	280	277	280	286	290	294	283	569	286
ALK(T)	178	178	181	181	181	181	181	182	181	18	163	159
LABPH	8.70	8.69	8.56	8.54	8.34	8.51	8.46	8.40	8.80	8.51	8.35	9.21
ONOS	391		376									
FECAL	410		<10	₽	\$	\$	8	8	50		×10	₽
댎	9.10	9.00	8.50	8.50	8.50	8.50	8.60	8.60	8.50	8.40	8.40	9.20
DISOX	16.8	17.0	13.3	13.5	13.5	13.5	11.8	12.0	16.0	9.0	8.0	13.0
SDISK	2.0	2.0	2.0									
ATEMP	40	40	51			38						
WTEMP	42	40	38	41	41	45	29	99	63	26	7	92
DEPTH	0.5	13.0	0.5	12.0	12.0	0.5	12.0	0.5	0.5	11.0	10.0	0.5
SAMPLE	URFACE	BOTTOM	SURFACE	BOTTOM	BOTTOM	URFACE	BOTTOM	URFACE	URFACE	BOTTOM	BOTTOM	URFACE
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11.KE	1045	1050	1130	1130	1130	1120	1045	1040	1000	1000	945	945
DATE	03/13/90	03/13/90	03/27/90	03/27/90	04/10/90	04/10/90	05/08/90	05/08/50	05/22/50	05/22/50	06/92/90	06/52/90

APPENDIX B Lake Basin Cross Sections

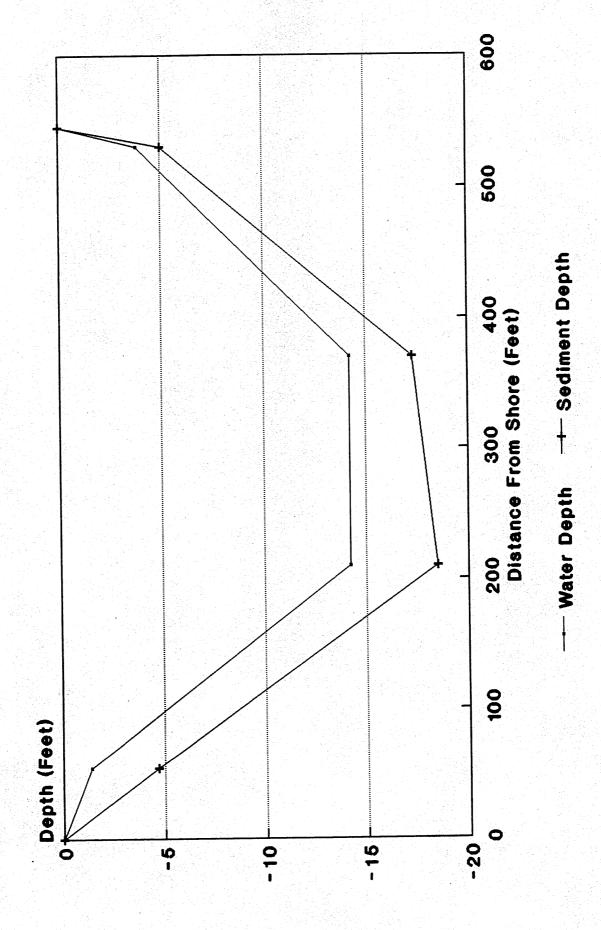


Note: Survey Data Collected 2/16/89.

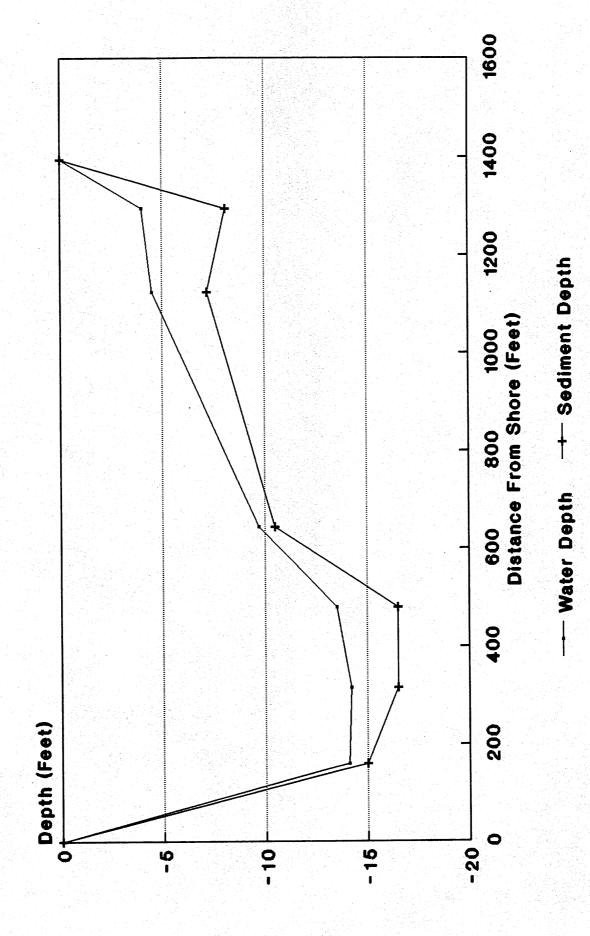
Burke Lake Cross-Section Data Cross-Section 1



Burke Lake Cross-Section Data Cross-Section 2

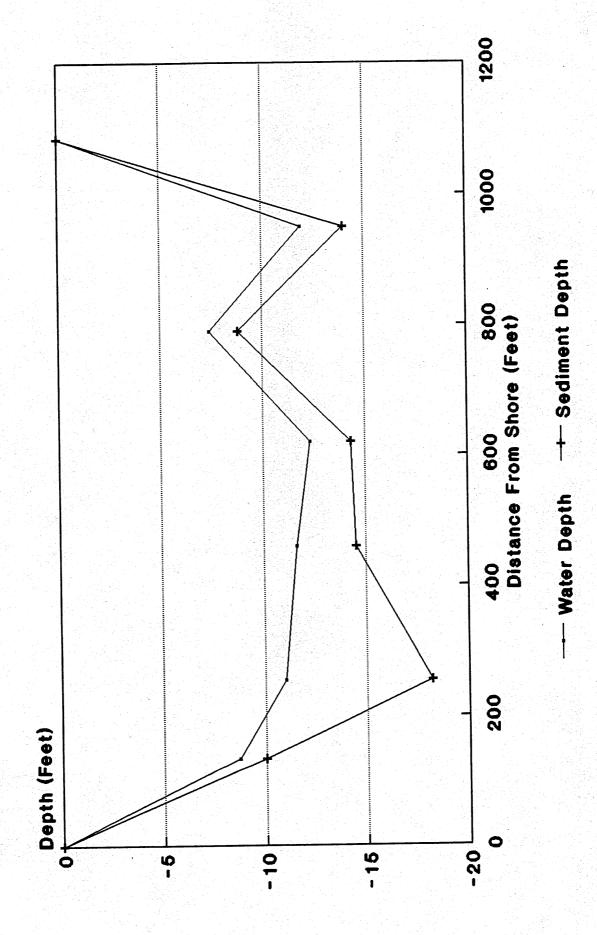


Burke Lake Cross-Section Data Cross-Section 3

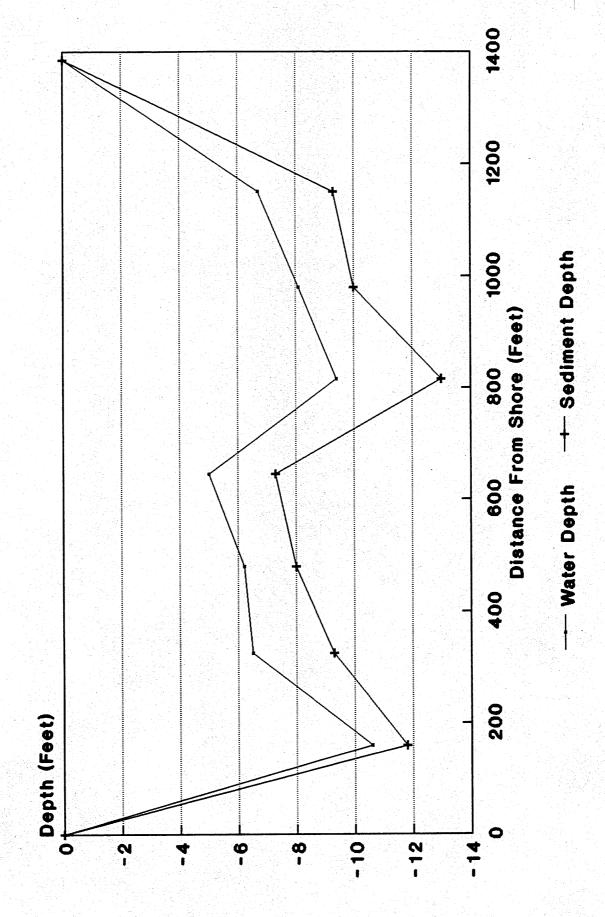


Data Collected 2/16/89

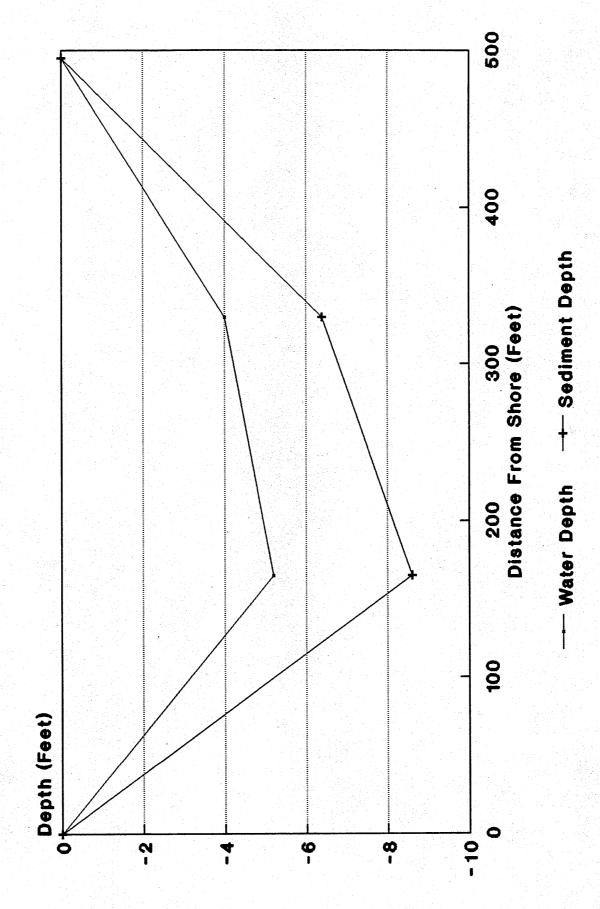
Burke Lake Cross-Section Data Cross-Section 4



Burke Lake Cross-Section Data Cross-Section 5

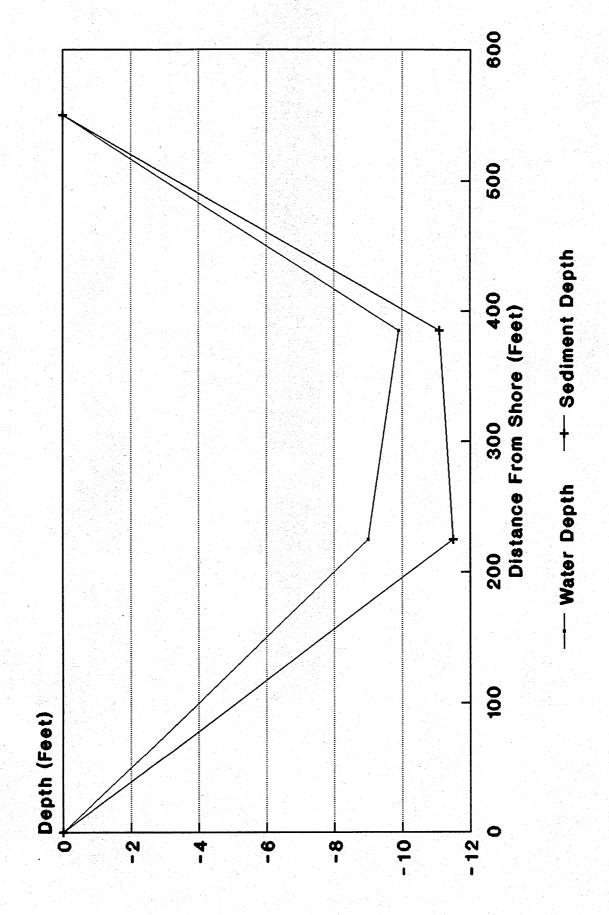


Burke Lake Cross-Section Data Cross-Section 6

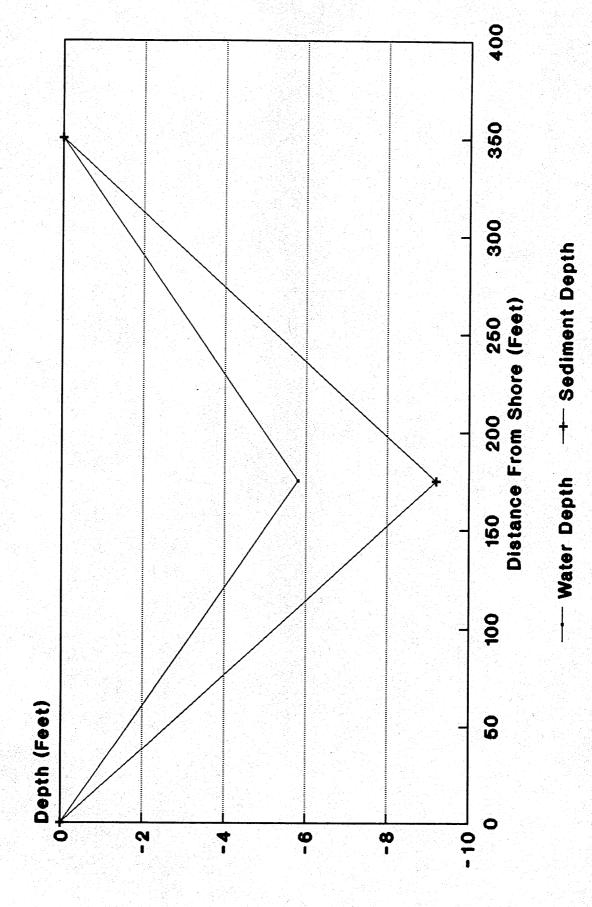


Data Collected 2/16/89

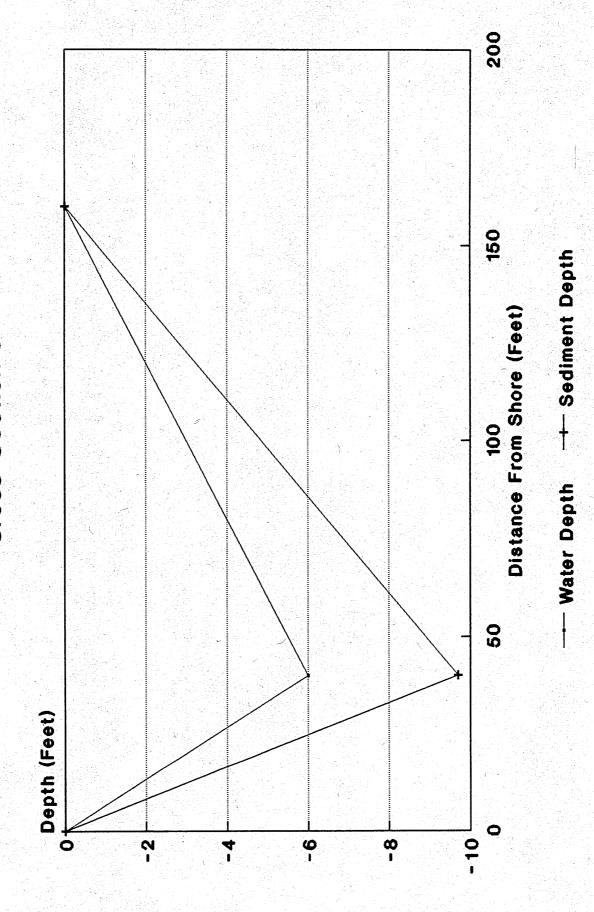
Burke Lake Cross-Section Data Cross-Section 7



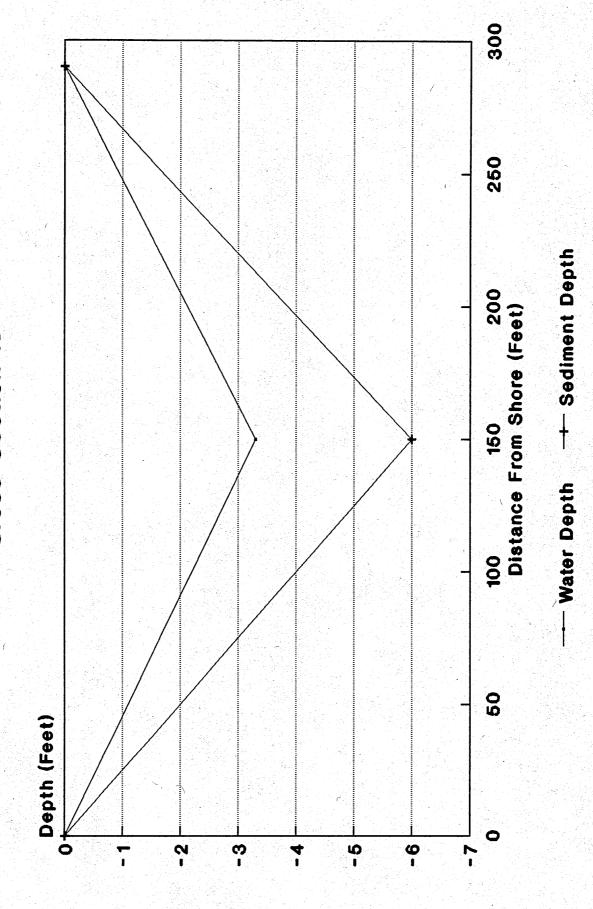
Burke Lake Cross-Section Data Cross-Section 8



Burke Lake Cross-Section Data Cross-Section 9



Burke Lake Cross-Section Data Cross-Section 10



APPENDIX C
Elutriate Sample Data

DEPARTMENT OF THE ARMY Missouri River Division, Corps of Engineers Division Laboratory Omaha, Nebraska

Project: South Dakota Department of Water and Natural Resource

Customer Sample Id: Burke Lake Date Sample Taken: 16 Feb 89 Lab Sample No: M-983

Date Sample Received: 21 Feb 89 Sample Description: Water and Sediment

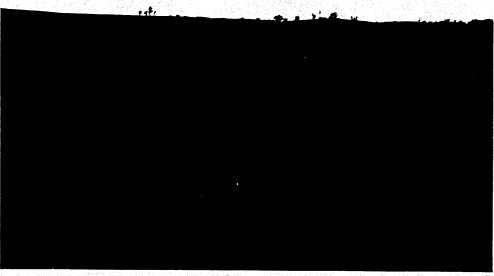
Sample Container: 2-1gal glass (water) and 1-1gal glass (sediment)

Time Sample Taken: 3:45pm

Comments: Burke Lake (Center of Lake, PT 12) Gregory County, SD

Analysis			Receiving		Elutriate	
	Sediment Result	Units	Water Result	Units	Water Result	Unit
	Kesutt		nesut c			
Ammonia Nitrogen			0.82	mg/L	3.6	mg/L
Chemical Oxygen Demand			43	mg/L	57.0	mg/L
Total Cyanide			<0.02	mg/L	<0.02	mg/L
Nitrate-Nitrite Nitrogen			0.28	mg/L	<0.01	mg/L
Total Phosphorus			0.22	mg/L	0.14	mg/L
Total Kjeldahl Nitrogen			5.3	mg/L	1.6	mg/L
Oil and Grease			<5	mg/L	10	mg/L
Antimony	0.42	mg/Kg	<1	ug/L	<1	ug/L
Arsenic	1.5	mg/Kg	2	ug/L	4	ug/L
Barium	340	mg/Kg	420	ug/L	210	ug/L
Beryllium	1.0	mg/Kg	<1	ug/L	<1	ug/L
Cadmium	<0.5	mg/Kg	<0.1	ug/L	<0.1	ug/L
Chromium	20	mg/Kg	<1	ug/L	(1)	ug/L
Copper	26	mg/Kg	<5	ug/L	<5	ug/L
Iron	27000	mg/Kg	70	ug/L	220	ug/L
	<5		<1	**************************************	<1	ug/L
Lead		mg/Kg	13	ug/L	10	
Magnesium	7400	mg/Kg		mg/L	990	mg/L
Manganese	630	mg/Kg	13	ug/L		ug/L
Mercury	<0.1	mg/Kg	<0.2	ug/L	<0.2	ug/L
Selenium	_1.2	mg/Kg	<1	ug/L		ug/L
Zinc	77	mg/Kg	<2	ug/L	<2	ug/L
Nickel	12	mg/Kg	2	ug/L	1	ug/L
Aluminum	33000	mg/Kg	<40	ug/L	<40	ug/L
Calcium	43000	mg/Kg	51	mg/L	42	mg/L
Sodium	240	mg/Kg	16	mg/L	16	mg/L
Potassium	4500	mg/Kg	24	mg/L	14	mg/L
Silver	~ 1 ~ .	mg/Kg	<0.2	ug/L	<0.2	ug/L
Simazine (Princep)	<100	ug/Kg	<0.1	ug/L	<0.1	ug/L
Metribuzin (Lexone)	<100	ug/Kg	<0.1	ug/L	<0.1	ug/L
Atrazine (Aatrex)	<100	ug/Kg	<0.1	ug/L	<0.1	ug/L
Aldrin	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
alpha-BHC	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
beta-BHC	<10	ug/Kg/	<0.01	ug/L	<0.01	ug/L
gamma-BHC (Lindane)	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
Hirex	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
Chlordane	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
P'P"DDD	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
P'P"DDE	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
P/PHDDT	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
Dieldrin	<10	ug/Kg	<0.01	ug/L	<0.01	Ug/L
	<10	ug/Kg	<0.01	Ug/L	<0.01	ug/L
Endosulfan I	<100		<0.1	Ug/L	<0.1	ug/L
Propachior (Ramrod)	<100	ug/Kg	<0.1	A STATE OF THE STA	<0.1	ug/L
Metolachlor (Dual)	<100	ug/Kg	<0.1	ug/L	<0.1	ug/L
Alachlor (Lasso)		ug/Kg	<0.1	ug/L	<0.1	ug/l
Diazinon	<100	ug/Kg		ug/L	<0.01	ug/l
Endrin	<10	ug/Kg	<0.01	ug/L	<0.01	
Heptachlor	<10	ug/Kg	<0.01	ug/L		ug/l
Heptachlor epoxide	<10	ug/Kg	<0.01	ug/L	<0.01	ug/L
Methoxychlor	<10	ug/Kg	<0.01	ug/L	<0.01	ug/l
Toxaphene	<500	ug/Kg	<0.5	ug/L	<0.5	ug/l
PCB-1016	<100	ug/Kg	<0.1	Ug/L	<0.1	ug/l
PCB-1221	<100	ug/Kg	<0.1	ug/L	<0.1	ug/l
PCB-1232	<100	ug/Kg	<0.1	ug/L	<0.1	ug/l
PCB-1242	<100	ug/Kg	<0.1	ug/L	<0.1	ug/l
PCB-1248	<100	ug/Kg	<0.1	ug/L	<0.1	ug/l
PCB-1254	<100	ug/Kg	<0.1	ug/L	<0.1	ug/l
PCB-1260	<100	ug/Kg	<0.1	ug/L	<0.1	ug/L

APPENDIX D
Watershed Photographs



Photograph 1. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking west towards abandoned

landfill.

Description: D/F study monitoring well #2. Well is located directly east and approximately 1500 feet downgradient of the abandoned city landfill.



Photograph 2. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking southwest towards the abandoned

landfill.

Description: D/F study monitoring well #3. Well is

located northeast and approximately 1750 feet downgradient of the abandoned city landfill.



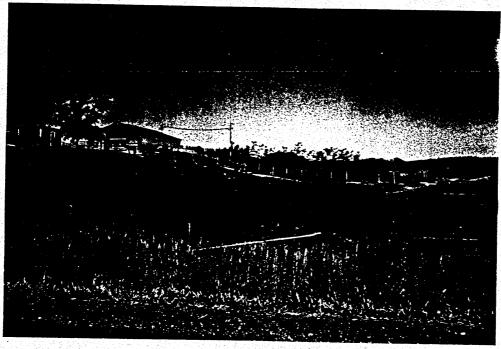
Photograph 3. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking northeast towards the western

tributary to Burke Lake.

Description: The tributary bed was dry at this time of year and was being utilized for grazing horses.



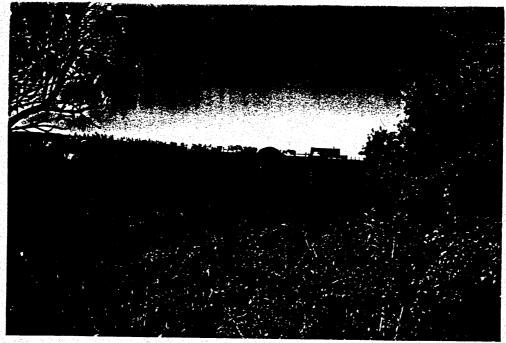
Photograph 4. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking north from the east-west county

road which leads to the lake access road.

Description: Manure piles in the watershed of Burke Lake. During precipitation events, this manure can wash into a tributary to Burke Lake and can impact groundwater.



Photograph 5. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking west from the southwest shore

of Burke Lake.

Description: Dairy operation located along the southwest and west borders of Burke Lake. Runoff from the feedlot can directly enter the lake.



Photograph 6. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking west from the southwest shore

of Burke Lake.

Description: Dairy operation located along the southwest and west borders of Burke Lake.



Photograph 7. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking east from the dairy operation.

Description: Aquatic vegetation on Burke Lake.



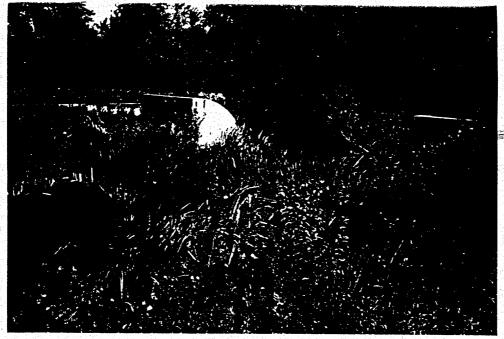
Photograph 8. Date: 10-3-90

Photographer: P. Szewczykowski

Orientation: Looking east at Burke Lake near

tributary sampling site #3.

Description: Turbidity in Burke Lake.



Photograph 9. Date: 10-3-90

Photographer: P. Szewczykowski Orientation: Looking northeast.

Description: The outlet structure at Burke Lake.

The structure is overgrown with vegetation.



Photograph 10. Date: 10-3-90

Orientation: Looking east and downstream from the

outlet.

Description: Debris and trees at the bottom of the

outlet.

OFFICE OF WATER RESOURCES MANAGEMENT

TIM BJORK - OFFICE ADMINISTRATOR

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JIM ANDERSON - ENGINEER
RICH HANSON - PROGRAM SCIENTIST
PHIL LIDEL - ENGINEER
KEN MADISON - SCIENTIST
ANDREW REPSYS - SCIENTIST
BILL STEWART - SCIENTIST
GENE STUEVEN - SCIENTIST
PAUL SZEWCZYKOWSKI - SCIENTIST